



Distribution of ^{137}Cs in the Bohai Sea, Yellow Sea and East China Sea: Sources, budgets and environmental implications

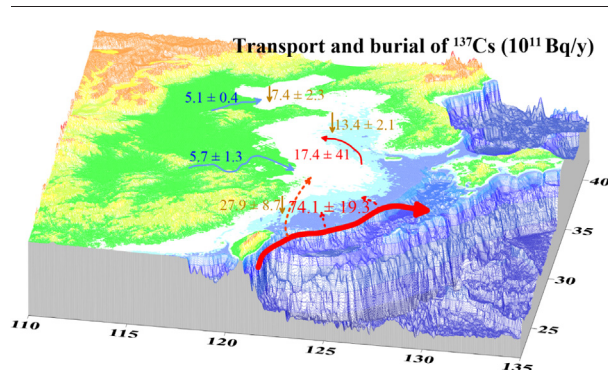
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HIGHLIGHTS

- The spatial distribution of ^{137}Cs in surface sediments and seawater was investigated.
- Oceanic inputs dominate the source of ^{137}Cs in eastern Chinese seas.
- The apparent half-lives of ^{137}Cs are 7.7–15.1 y in eastern Chinese seas.
- ^{137}Cs is useful to track fine sediment transport in river-dominated marginal seas.
- ^{137}Cs is useful to trace water mass movement and exchange in marginal seas.

GRAPHICAL ABSTRACT



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ABSTRACT

The transport processes and fate of ^{137}Cs in eastern Chinese seas (ECSs) that consists of the Bohai, Yellow, and East China Seas, have not been well established. In this study, we measured the concentrations of ^{137}Cs in the ECSs water and surface sediments during 2013–2014. Combined with a number of published ^{137}Cs inventory data from drainage basins and sediment accumulation rates in the ECSs, the distribution, sources and budgets of ^{137}Cs in the ECSs were investigated. The ^{137}Cs activity in the water column and surface sediments ranged from 0.03 to 1.92 Bq/m³ and from 0.30 to 5.22 Bq/kg, respectively. No ^{134}Cs signal was observed, suggesting that the Fukushima accident had limited impact on the ECSs during the investigation period. Mass balance of ^{137}Cs suggests that at least 7.4×10^{12} Bq/y of ^{137}Cs imported into the ECSs from the Northwestern Pacific that accounts for 0.7% of the ^{137}Cs transported by the Kuroshio Current, and this value is 5.2 times higher than the sum of atmospheric fallout and total riverine input. The apparent half-lives of ^{137}Cs are estimated to be 15.1 y for the ECS and 7.7 y for the YS. The vertical profiles in the continental shelf edge and the Yangtze River Estuary reveal that the upwelling of the Kuroshio Subsurface Water is the main mechanism of ^{137}Cs import into the ECSs. The high level of ^{137}Cs in oceanic water masses and the low level of ^{137}Cs in riverine and coastal waters make ^{137}Cs a good indicator for tracing water mass movement and interaction. In addition, good correlation between ^{137}Cs activity and mean grain size (φ) indicates that ^{137}Cs can serve as an effective tracer to track dispersal pathways of fine sediments in river-dominated marginal seas.

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1. Introduction

Anthropogenic radionuclides have been introduced to marine environments by atmospheric deposition as a consequence of atmospheric nuclear weapons testing, nuclear accidents, and discharges from nuclear installations. Among these radionuclides, ^{137}Cs has received the most attention not only due to its long half-life (30.17 years) and high radiotoxicity, but also for its implication as a powerful tool for oceanography. Considering the well-defined origin and conservative behavior of ^{137}Cs in the open ocean, it is useful to trace water mass movement, mixing, and circulation (Bowen et al., 1980; Aarkrog et al., 1983; Cochran et al., 1987; Aoyama et al., 2011). Moreover, vertical distribution of ^{137}Cs in sediment cores is widely utilized for reconstruction of sedimentary deposition history (Fuller et al., 1999; Su and Huh, 2002; Pfitzner et al., 2004; Wang et al., 2016).

The Pacific Ocean is recognized as a major recipient of ^{137}Cs not only because of its sheer size, but also due to the presence of the North Pacific Proving Grounds (NPPG) in the Marshall Islands (Povinec et al., 2004). ^{137}Cs in the Pacific Ocean has two major sources: global fallout and close-in fallout from the NPPG (Bowen et al., 1980; Yamada et al., 2006). Aarkrog (2003) has estimated that the ^{137}Cs from global fallout to the Pacific was approximately 311 PBq, while the close-in fallout of ^{137}Cs from the NPPG may be on the order of 100 PBq. In addition, the deposition rates of ^{137}Cs from global fallout varied by latitude, and the highest deposition occurred in the mid-latitudes of the Northwestern Pacific (WNP) (Aoyama et al., 2006). More recently, the Fukushima accident caused a large amount of ^{137}Cs to be released, and most estimates suggested that the ^{137}Cs released into the WNP ranged from 15 PBq to 20 PBq (Tsubono et al., 2016; Aoyama et al., 2016; Buesseler et al., 2017). Therefore, the Pacific Ocean, especially the WNP, has always been a hotspot of ^{137}Cs study. The distribution, transport processes and fate of ^{137}Cs in the WNP have already been intensively investigated in a series of studies (e.g. Bowen et al., 1980; Aoyama and Hirose, 1995; Povinec et al., 2003; Yamada et al., 2006), but have not been well discussed in its marginal seas—the eastern Chinese seas (ECSs) that include the Bohai Sea (BS), Yellow Sea (YS) and East China Sea (ECS), only some outdated and low resolution data have been reported (Nagaya and Nakamura, 1992; Wu et al., 2013). The ECSs link the world's largest continent (Asia–Europe) and ocean (Pacific Ocean). Considering the conservative feature of ^{137}Cs in seawater, it can be entrained in the northward flowing Kuroshio Current (KC), a branch of the North Equatorial Current (NEC), and then spread widely to the WNP and its marginal seas. This long-range transport process has been confirmed by $^{239+240}\text{Pu}$ and their atom ratios (Yamada et al., 2006; Liu et al., 2011; Wang et al., 2017a). By comparing concentrations of ^{137}Cs in the coastal seawater of Japan, Kasamatsu and Inatomi (1998) pointed out that surface water with a higher ^{137}Cs concentration is transported from the WNP via the Tsushima Current (TC), a branch of the KC. The similar ^{137}Cs activities in the Japan/East Sea and WNP also suggested that ^{137}Cs from the NPPG can be transported to the Japan/East Sea (Ito et al., 2003; Povinec et al., 2004). The ECSs are located between the Japan/East sea and the WNP; thus, it is reasonable to believe that ^{137}Cs from the WNP can be transported into the ECSs. However, a study of the behavior and fate of ^{137}Cs in the ECSs is still limited.

The ECSs are typical river-dominated marginal seas. The Yangtze River and Yellow River, the two largest rivers on the Eurasian continent, discharge ~1.6 billion tons of sediments annually to the ECSs, accounting for 10% of the world's annual sediment discharge (Milliman and Farnsworth, 2013). These sediments can carry many terrestrial particles that can bond with ^{137}Cs , which could affect the distribution and inventory of the ^{137}Cs in the ECSs. Most importantly, the enormous riverine sediment discharge leads to a high concentration of suspended particles in the ECSs. Meanwhile, the current system in the ECSs is complex and the water depth is shallow, and therefore, the behavior of the ^{137}Cs in the ECSs is believed to be quite different from the open ocean. In general, the scavenging and burying processes by association with the

particles could be more intensive than those in the open ocean (Carpenter et al., 1987; Nagaya and Nakamura, 1992). In addition, boundary scavenging processes could occur in the highly turbid water column of the nearshore areas and enhance the import of ^{137}Cs from offshore areas (Du et al., 2010). These characteristics make the transport processes of ^{137}Cs in the ECSs complicated.

In our previous work, the results showed that oceanic input is a dominant source of ^{137}Cs in the ECS (Du et al., 2010; Zhao et al., 2018). Now we extend our research to the ECSs and present more comprehensive data to do the following: (i) describe the distribution characteristics of ^{137}Cs in the water column and surface sediments in the ECSs; (ii) construct the ^{137}Cs budget in the ECSs within a source-sink perspective by calculating the atmospheric fallout, riverine input, oceanic input, natural decay, and burial flux; (iii) evaluate the tracer application of ^{137}Cs in marine environments. We quantify the interlinked budget balance of ^{137}Cs between the BS, YS and ECS. This work will contribute to our understanding of not only the transport processes and ultimate fate of ^{137}Cs but also the biogeochemical cycles of other analog materials in river-dominated marginal seas.

2. Materials and methods

2.1. Study area

The ECSs are surrounded by mainland China, the Korean Peninsula, Taiwan Island and Ryukyu Island. There are many rivers flowing into the ECSs including the Yangtze River and Yellow River, two of the largest river systems in the world (Fig. 1). The BS is a Chinese inland sea with an average water depth of 18 m, connected with the YS by the Bohai Strait. The YS is a typical semienclosed epicontinental sea between the BS and ECS, with an average water depth of 44 m. The YS is separated from the ECS to the south by an arbitrary line connecting the north of the Yangtze River mouth with Cheju Island. The ECS is a very broad and flat continental shelf sea, with a maximum width of 640 km. Based on the water depth, it can be divided into an inner-shelf area, outer-shelf area and slope, and Okinawa Trough area, with water depths below 70 m, between 70 m and 200 m and above 200 m, respectively (Su and Huh, 2002). The ECS, together with the YS and BS, constitute a series of interlinked marginal seas in the WNP.

Due to effect of monsoon, Kuroshio Current, fluvial runoff and tidal motion, the circulation pattern in the ECSs is very complex and dynamic (Fig. 1), as described in Su (2001). In general, there are two current systems of which one is warm and salty current system of oceanic origin, and another one is cold and diluted coastal current system (Guan, 1994). The former moves northward including the Kuroshio Current (KC) along the edge of the ECS continental shelf and the Taiwan Warm Current (TWC) from the Taiwan Strait. The Yellow Sea Warm Current (YSWC) is a branch of KC and flows northwest into the YS, and its extension can reach the BS through northern Bohai Strait. As for coastal current system, most of coastal currents flow southward including the Bohai Sea Coastal Current (BSCC), Yellow Sea Coastal Current (YSCC) and West Korea Coastal Current (WKCC), while the Zhejiang-Fujian Coastal Current (ZFCC) flows northward in summer and southward in winter owing to the change of prevailing monsoon. The Changjiang Diluted Water (CDW) flows toward northeast in flood season and turns to south in dry season.

2.2. Sampling and analysis

Two cruises onboard the R/V “Dongfanghong 2” were conducted in the ECS during 6 to 28 August 2013 and in the YS-BS during 28 April to 17 May 2014. A total of 107 seawater samples (including 83 surface samples and 24 deep layer samples) and 148 surface sediment samples (0–1 cm) were collected (Fig. 1). Surface seawater samples (0–0.5 m) were collected by a submersible pump and deep layer samples were collected by using a CTD rosette system. All water samples filtered

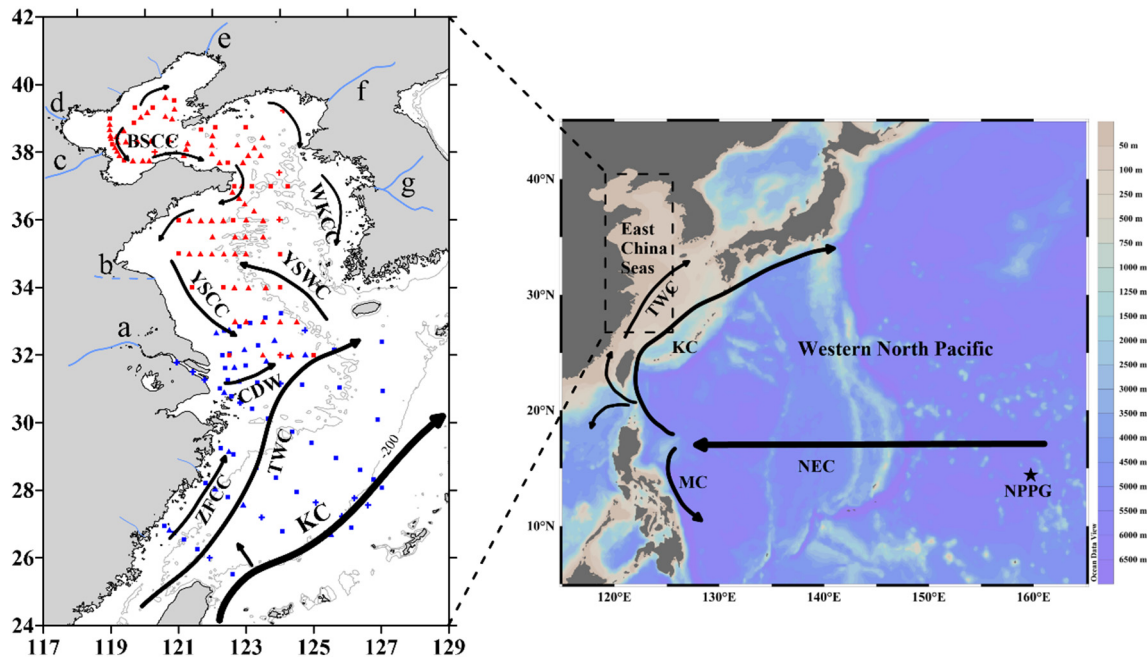


Fig. 1. Map showing the sampling locations and circulation pattern in the Eastern Chinese Seas. Sampling stations in August 2013 and May 2014 are marked in blue color and red color, respectively. Triangle symbols represent that only surface sediments were collected at that station. Cross symbols represent only seawater samples were collected. Square symbols represent both sediment and seawater samples were collected. Regional ocean circulation pattern is modified from Su (2001) and Hu et al. (2015). BSCC: Bohai Sea Coastal Current; YSCC: Yellow Sea Coastal Current; YSWC: Yellow Sea Warm Current; WKCC: West Korea Coastal Current; CDW: Changjiang Diluted Water; ZFCW: Zhejiang-Fujian Coastal water; TWC: Taiwan Warm Current; KC: Kuroshio Current; NEC: North Equatorial Current; MC: Mindanao Current. Several large rivers are also plotted, including a: Yangtze River, b: old Yellow River, c: Yellow River, d: Haihe, e: Liaohe, f: Yalujiang, g: Han River.

through a pre-cleaned special polypropylene cartridge with 0.5 μm pore-size (Wang et al., 2017b). Sediment samples were collected using a box sampler, then placed in polyethylene bags and frozen in a refrigerator for further analysis.

^{137}Cs activities in seawater samples were measured following the method described by Zhao et al. (2018). Briefly, filtered seawater samples (~100 L each) were acidified to $\text{pH} < 2$ with concentrated HCl, then added 25 mg CsCl as carrier and chemical yield tracer. After equilibration, 40 g ammonium phosphomolybdate (AMP) was added and stirred for 30 min until well mixed. After settling about 24 h, the supernatants were siphoned out, then the residual was transferred into a 2 L plastic bottle and brought back to the laboratory. In lab, the residual was centrifuged and freeze-dried for three days to separate AMP-Cs precipitate. The dried sample was homogenized and sealed in a plastic tube. Finally, the activities of ^{137}Cs in prepared samples were measured using a well-type HPGe gamma spectrometer (GWL-120210-S, ORTEC). ^{137}Cs was detected at 661.6 keV peak with the branch ratio of 85.1%. In addition, considering the Fukushima accident, ^{134}Cs was also concerned during the measurement, however, there was no signal at 604 keV in gamma spectrums. The mean chemical recovery was $77.8\% \pm 3.2\%$, obtained by using the atomic absorption spectrophotometer (AA-800, Perkin-Elmer) to determine stable Cs in the AMP-Cs precipitate. The IAEA standard source (IAEA-447) was analyzed for quality control, and the result agreed with the certificated value well.

The sediment samples were oven dried at 60 $^{\circ}\text{C}$ for several days. An aliquot of dry sediments was used to determine the grain size. Briefly, 0.1–0.2 g sediments were placed in beakers, then added 5 mL 10% H_2O_2 and 5 mL 0.2 M HCl to remove organic matter and coatings of metal oxides. After about 24 h, 10 mL 0.5 M sodium hexametaphosphate solution was added and oscillated in ultrasonic bath for 10 min to make the sediments disaggregate completely, then analyzed by using a laser particle analyzer (LS100Q, Beckman, USA). Determination of ^{137}Cs activities in sediments followed the method described by Du et al. (2010). The dried sample was ground,

homogenized, and sealed in a plastic box (70 mm diameter \times 70 mm height), then measured by a HPGe γ -ray detector (ADC-1000-I, ORTEC) with a relative counting efficiency of 35% and an energy resolution of 1.8 keV (at 1332 keV). Data reported for ^{137}Cs in all samples were decay-corrected to sampling date.

3. Results

3.1. Spatial distribution of ^{137}Cs in ECSs' surface water

The temperature, salinity and ^{137}Cs activity in the ECSs' waters are listed in the Supplemental file (Table S1). The spatial distribution of ^{137}Cs activity in the surface waters of the ECSs are shown in Fig. 2. In the ECS, ^{137}Cs activity ranged between 0.12 and 1.59 Bq/m^3 (mean: $0.62 \pm 0.32 \text{ Bq}/\text{m}^3$, $n = 53$). The highest ^{137}Cs concentration was observed in the shelf edge area of the ECS, while the lowest was found in the Yangtze Estuary and extended to the northeast to Cheju island (Fig. 2a). Our results were lower than the values observed in the Northwest Pacific (1.0–2.2 Bq/m^3 , mean: $1.6 \pm 0.4 \text{ Bq}/\text{m}^3$, decay-corrected to 2014) (Povinec et al., 2003), East China Sea (0.70–1.33 Bq/m^3 , mean: $1.05 \pm 0.07 \text{ Bq}/\text{m}^3$, decay-corrected to 2014) (Wu et al., 2013), South China Sea (0.99–1.41 Bq/m^3 , mean: $1.12 \pm 0.14 \text{ Bq}/\text{m}^3$) (Wu, 2018) and Luzon Strait (0.43–2.44 Bq/m^3 , mean: $1.04 \pm 0.13 \text{ Bq}/\text{m}^3$) (Zhou et al., 2018). ^{137}Cs concentrations in the surface waters of the YS and BS were generally low, ranging 0.08–0.68 Bq/m^3 (mean: $0.38 \pm 0.17 \text{ Bq}/\text{m}^3$, $n = 21$) and 0.03–0.09 Bq/m^3 (mean: $0.05 \pm 0.02 \text{ Bq}/\text{m}^3$, $n = 9$), respectively (Fig. 2b). The southern region of the YS had higher ^{137}Cs activities, with a sharply decreasing trend northward, but smoothly westward. The ^{137}Cs activities in the YS were much lower than results measured in 1987 (2.93–4.56 Bq/m^3 , mean: $3.80 \pm 0.70 \text{ Bq}/\text{m}^3$) (Nagaya and Nakamura, 1992) and in 1999–2000 (1.62–3.41 Bq/m^3 , mean: $2.7 \pm 0.2 \text{ Bq}/\text{m}^3$) (Hong et al., 2006), showing a decreasing trend over time.

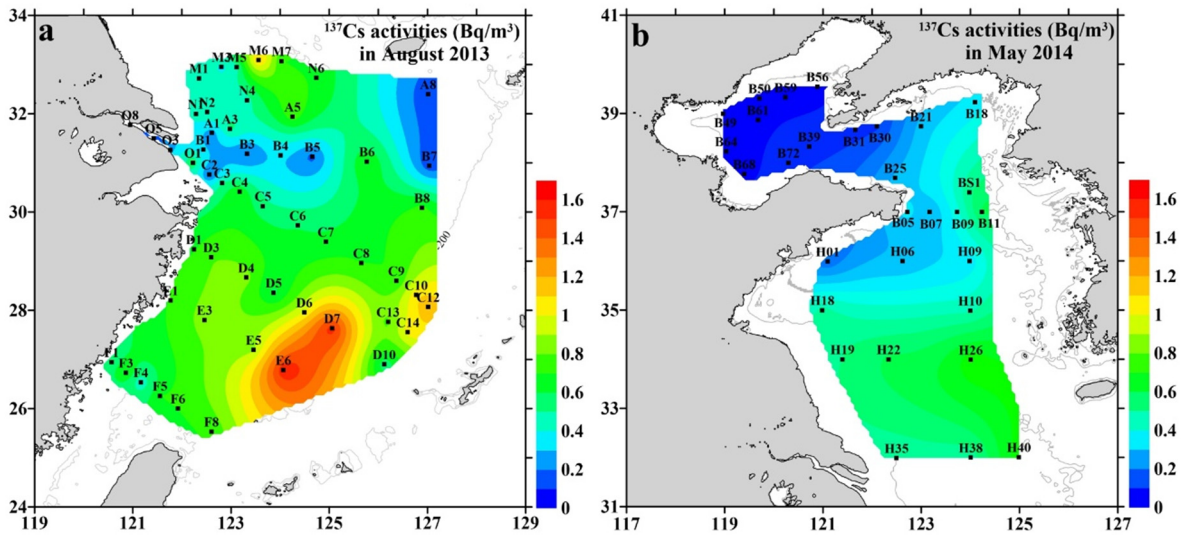


Fig. 2. Spatial distribution of ^{137}Cs activities in the surface water of the ECS (a) and YS-BS (b).

3.2. Vertical profiles of ^{137}Cs in the water column

Fig. 3 shows the vertical profiles of ^{137}Cs activities in the ECSs. At a few stations on the ECS shelf edge (e.g., F8, D10, C12 and C13), ^{137}Cs activity maxima were observed in the subsurface layer at a depth of 100–200 m (Fig. 3a), although the sampling depth did not cover the entire water column. In fact, similar ^{137}Cs profiles have been observed in the subtropical and equatorial North Pacific, with ^{137}Cs peaking at approximately 200 m depth (Aoyama et al., 2008), and in the Luzon Strait, with the maximum ^{137}Cs level occurring at 150 m depth (Zhou et al., 2018). In the outer shelf region, ^{137}Cs activity decreased with increasing depth at station E6, while in the estuary area, a reverse trend was observed at station C2 (Fig. 3b). At all stations in the YS, ^{137}Cs concentrations in the bottom layer were slightly lower than in the surface layer (Fig. 3c).

3.3. Grain size distribution of surface sediments

The mean grain size (φ , $\varphi = -\log_2^M$, M refers to mean grain size in mm) and percentage of clay component of surface sediments are listed in Table S2 and plotted in Fig. 4a and b. The mean grain size (φ) ranged from 2.0 to 7.9 and clay content ranged from 3.3% to 48.6%. In the ECS,

sediment pattern is characterized by “fine-coarse-fine” feature from the inner shelf to shelf slope. In the Zhe-Min coast and the ECS shelf slope area, the clay contents were higher than 25% and sand contents were lower than 10%, while in the outer shelf region the sand content was larger than 60% and clay content was lower than 15%. In the southwestern Cheju island, the clay content was also higher than 25%. The highest clay content was found in the central YS with percentage >40%. The sand contents in this area were generally lower than 5%. In the north of the Yangtze estuary and the western part of the Southern YS, two small patches of coarse sediments composed of clay (<10%), silt (<25%) and sand (>65%) were observed. (See Fig. 4b.)

3.4. Spatial distribution of ^{137}Cs in surface sediments

The ^{137}Cs activity in surface sediments is listed in Table S2 and plotted in Fig. 4c. The ^{137}Cs activities in surface sediments of the ECS, YS and BS ranged between 0.30 and 5.22 Bq/kg (mean: 1.45 ± 1.00 Bq/kg), 0.54 and 5.02 Bq/kg (mean: 1.81 ± 1.09 Bq/kg) and 0.52 and 4.58 Bq/kg (mean: 1.57 ± 0.92 Bq/kg), respectively. In the ECS, ^{137}Cs activities varied by more than an order of magnitude. The highest values were observed in the Zhe-Min coastal area with a sharply decreasing trend offshore. The lowest activities were found in the outer shelf. Activities then increased slightly near the shelf slope area. In the central YS, a

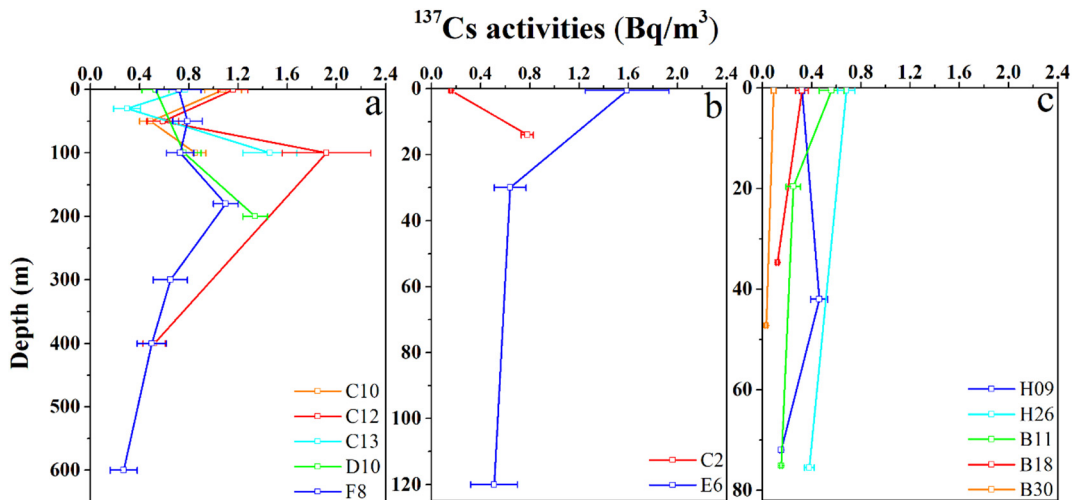


Fig. 3. Vertical profiles of ^{137}Cs activities in the water column of the ECSs.

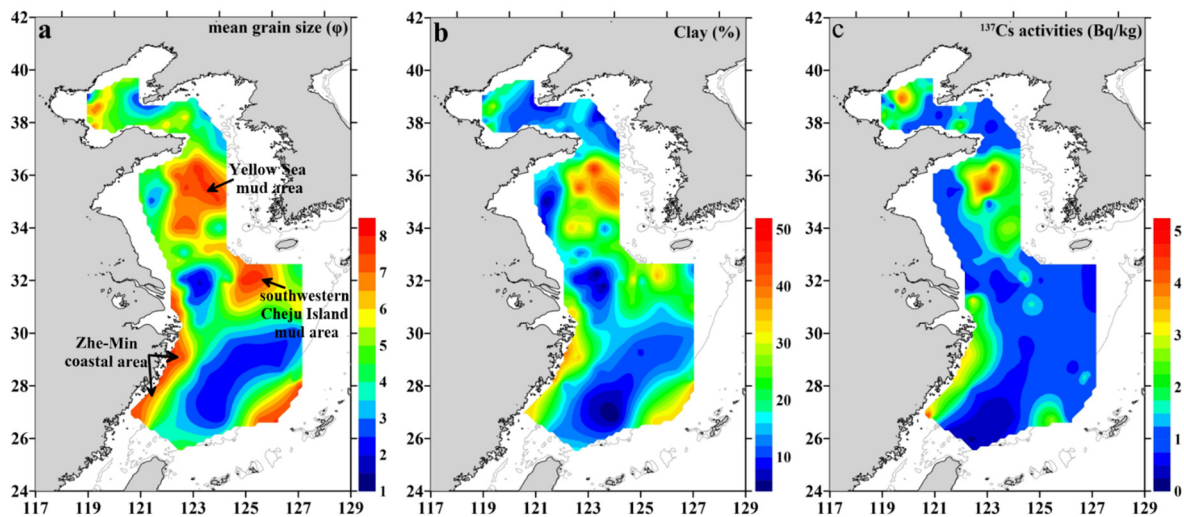


Fig. 4. Spatial distribution of mean grain size (ϕ), clay component (%) and ^{137}Cs activities (Bq/kg) in surface sediments of the ECSs.

large patch of high activity was observed. The high value region was also found in the central BS. In the western part of the YS and the BS strait, ^{137}Cs activities in surface sediments were very low. In general, the spatial distribution of ^{137}Cs activities in surface sediments was similar to the mean grain size pattern.

4. Discussion

4.1. Sources and sinks of ^{137}Cs in the ECSs

The major sources of ^{137}Cs in the ECSs are potentially from atmospheric fallout from global weapons testing, riverine input from the soil erosion, emission from the coastal nuclear power plants (NPPs), influence of the Fukushima accident and oceanic input from the WNP. The output of ^{137}Cs in the ECSs is burial to the seafloor sediments and natural decay. So far, there was no any report on the leakage of radionuclide contaminants from Chinese coastal NPPs, and sediment cores in the ECSs did not record the signal of such input of ^{137}Cs (Huh and Su, 1999; Li et al., 2013; Zhu et al., 2014; Liu et al., 2015). Besides, ^{137}Cs in coastal water of the ECSs was lower than offshore area, this observation further confirmed that the emission from the coastal NPPs could be negligible. The majority of the Fukushima accident-derived ^{137}Cs has been proved to be transported eastward (Buesseler et al., 2017), and in gamma spectrums of all samples we did not observe the signal of ^{134}Cs at 604 keV, thus the influence of the Fukushima accident could also be negligible.

4.1.1. Input from atmospheric fallout of ^{137}Cs

As a typical artificial radionuclide, the atmospheric fallout of ^{137}Cs is in a nonsteady state. Large-scale atmospheric nuclear weapons testing was conducted in 1961–1962, and the stratospheric residence time of ^{137}Cs was approximately 1 year. Accordingly, the annual deposition rate of ^{137}Cs peaked in 1963 (Hirose et al., 2008). Then, the deposition rate declined over time, except in 1986–1987 that was influenced by the Chernobyl accident. However, long-term observation in Tsukuba ($36^{\circ}03'\text{N}$, $140^{\circ}08'\text{E}$) showed that the declining trend of the ^{137}Cs deposition slowed or almost ceased since the 1990s, suggesting that current ^{137}Cs deposition is boosted by the resuspension and transportation of ^{137}Cs deposited on the East Asian continent (Hirose et al., 2008). Since our study area is also located in the Asian yellow dust plume, it is expected that the current deposition status over the ECSs is similar to the observation in Tsukuba. Therefore, we use the mean value ($0.22 \pm 0.09 \text{ Bq/m}^2/\text{y}$) measured in Tsukuba during 1990–2010 as the current ^{137}Cs deposition rate in the ECSs. This value is comparable to the result ($0.33 \pm 0.20 \text{ Bq/m}^2/\text{y}$) obtained in Shanghai ($31^{\circ}13'39''\text{N}$, $121^{\circ}23'56''$

E) during 2006–2007 (Du et al., 2010). The surface areas of our study areas in the BS, YS, and ECS were estimated by Google Earth to be $0.33 \times 10^5 \text{ km}^2$, $1.95 \times 10^5 \text{ km}^2$ and $3.82 \times 10^5 \text{ km}^2$, respectively. Multiplying by the ^{137}Cs deposition rate, the input of ^{137}Cs to the three seas by atmospheric fallout are calculated to be $(0.7 \pm 0.3) \times 10^{10} \text{ Bq/y}$, $(4.3 \pm 1.8) \times 10^{10} \text{ Bq/y}$ and $(8.4 \pm 3.4) \times 10^{10} \text{ Bq/y}$, respectively.

4.1.2. Riverine input from the soil erosion in the catchment

There are many rivers around the ECSs, 40 of which are $>1000 \text{ km}^2$ in catchment area. These rivers annually discharge tremendous amounts of terrestrial sediments into the ECSs and could be a potential source for the ^{137}Cs budget as a result of soil erosion in the drainage basin. Ideally, calculating a riverine contribution precisely should be based on field measurements in each river mouth. However, it is too massive of a project to carry out, and very few data of ^{137}Cs in river water have been published. Previous studies suggested that the riverine input of radionuclides can be roughly estimated by the following equation (modified from: Ravichandran et al., 1995; Baskaran et al., 1997):

$$I_d = A_d \times I_f \times f_e \quad (1)$$

where A_d is the area of the drainage basin, I_f is the reference inventory of radionuclide in soils (Bq m^{-2}), and f_e is the fraction of radionuclide eroded each year from drainage basin ($f_e = \ln 2/\text{residence time of radionuclide in drainage basin}$). This method has been applied successfully in the ECS to estimate the input of ^{210}Pb and $^{239+240}\text{Pu}$ derived from the Yangtze River (Su and Huh, 2002; Wang et al., 2017a).

As shown in the equation, the riverine input contribution depends on the catchment area and residence time of the radionuclide in the drainage basin. The residence time of ^{137}Cs in several different scale catchments has been investigated, with the value reported to be 800 y in the Rhone River (catchment area: $0.52 \times 10^4 \text{ km}^2$), 1500 y in the Saguenay River ($7.8 \times 10^4 \text{ km}^2$) and 4100 y in the Columbia River ($68 \times 10^4 \text{ km}^2$) (Smith and Ellis, 1982; Dominik et al., 1987; Beasley and Jennings, 1984). If the catchment area is larger, the corresponding residence time is longer. Therefore, the ^{137}Cs input from small rivers may not be neglectable. Considering massive catchment areas and the construction of many dams (e.g., the Three Gorges Dam and the Xiaolangdi Dam), we use the largest value of 4100 y as the residence time of ^{137}Cs in the Yangtze River and the Yellow River. For other small rivers around the ECSs, the ^{137}Cs residence time assignment depends on their catchment areas (Table S3). ^{137}Cs has been widely applied in determining soil erosion rates in mainland China and the Korean Peninsula (e.g. Zhang et al., 2003; Zhang et al., 2011; Meusburger et al., 2013; Ma et al., 2016; Li et al., 2017). As a consequence, there is an available

comprehensive dataset of ^{137}Cs reference inventories in soils (Table S4), which enables us to better estimate the riverine input of ^{137}Cs . The ^{137}Cs inventories in high latitudes (30°N – 40°N) are generally higher than those in low latitudes (20° – 30°N), which is consistent with global fallout patterns. To calculate riverine inputs, we divide the ^{137}Cs inventory data into four parts, and the corresponding arithmetic mean value in each part is shown in Fig. 5. The mean ^{137}Cs inventory in the Yellow River basin is $1594 \pm 251 \text{ Bq/m}^2$, and accordingly, the input of ^{137}Cs from the Yellow River is estimated to be $(2.0 \pm 0.3) \times 10^{11} \text{ Bq/y}$. The Yellow River discharge is approximately $0.7 \times 10^8 \text{ tons/y}$ of terrestrial sediments into the BS (CWRC, 2014). Thus, the mean ^{137}Cs activity in suspended particulates eroded from the drainage basin can be calculated to be $2.9 \pm 0.5 \text{ Bq/kg}$. This value is close to the activity of surface sediments near the Yellow River mouth. The Yangtze River delivers the most terrestrial ^{137}Cs into the ECS. The mean ^{137}Cs inventory in the Yangtze catchment is $1462 \pm 470 \text{ Bq/m}^2$. Thus, the riverine input is calculated to be $(4.5 \pm 1.4) \times 10^{11} \text{ Bq/y}$. However, this estimation is eight times larger than the filed measured value ($0.51 \times 10^{11} \text{ Bq/y}$) at the Xuliujing station (Du et al., 2010). The discrepancy may come from two aspects. First, the field measurement was carried out in 2006–2007, when the Yangtze basin runoff was at its lowest level in the last half century due to extremely low annual rainfall, high temperatures and the impounding of Three Gorges Dam from 135 m to 156 m (Dai et al., 2008). The global fallout of ^{137}Cs should be much higher in the 1960s compared to 2006–2007, which lead to a much higher riverine discharge of ^{137}Cs in the 1960s. Thus, it is not surprising that our estimated result represents a mean value between 1952 and 2013 and is higher than the measured value during 2006–2007. Second, the residence time of ^{137}Cs in the Yangtze basin is a highly uncertain factor, even though we used the largest value of the range (4100 y), the residence time might still be underestimated and the input of ^{137}Cs might be overestimated. Taking all rivers into the calculation (Table S3), the total riverine inputs to the BS, YS and ECS were estimated to be $(5.1 \pm 0.4) \times 10^{11} \text{ Bq/y}$, $(2.0 \pm 0.2) \times 10^{11} \text{ Bq/y}$ and $(5.7 \pm 1.3) \times 10^{11} \text{ Bq/y}$, respectively.

4.1.3. Sediment accumulation rates and ^{137}Cs burial

In the past few decades, the sediment accumulation rate estimated by radioisotopes (^{210}Pb , ^{137}Cs and $^{239+240}\text{Pu}$), as an important factor in sedimentary process studies in the ECSs, has been intensely investigated (e.g. DeMaster et al., 1985; Alexander et al., 1991; Li et al., 1993, 2006; Huh and Su, 1999; Su and Huh, 2002; Oguri et al., 2003; Fang et al., 2007; Li et al., 2013; Qiao et al., 2017). The highest sediment accumulation rates ($>2 \text{ cm/y}$) are observed near the Yangtze River mouth and the Yellow River mouth (Fig. 6a). In the coastal and inner shelf of the ECS, sediment accumulation rates are high but decrease eastward sharply to $<0.1 \text{ cm/y}$. Well-preserved sedimentary structures in the coast and inner shelf suggest that sediment deposition is dominant compared to other factors, even though remobilization and resuspension of sediments are active (McKee et al., 1983; DeMaster et al., 1985; Wang et al., 2016). Moreover, sediment accumulation rates derived from different radioisotopes showed a high consistency, indicating that biological perturbation is limited (DeMaster et al., 1985; Su and Huh, 2002). The sediment accumulation rates in the YS are generally low, with a decreasing trend from west to east. Relatively high rates in the western part are a result of sediment input eroded from the old Yellow River delta (Alexander et al., 1991). In the modern Yellow River delta front region, the accumulation rates decline from above 2 cm/y to below 0.2 cm/y with a sharp gradient.

At the same interpolation gridding density, by combining ^{137}Cs activities in surface sediments, bulk density and sediment accumulation rates, the spatial distribution of ^{137}Cs burial flux in surface sediments in the ECSs can be obtained (Fig. 6b). There are two patches of high ^{137}Cs burial flux. The first patch of high ^{137}Cs burial flux is located in the coastal area along the sediment dispersal path of the Yangtze River, with a sharply decreasing trend offshore. Such a narrow band with a high ^{137}Cs burial flux reveals that most of the terrestrial ^{137}Cs was carried by the Yangtze River deposits in the estuary and inner shelf area. The second patch of high ^{137}Cs flux is found in the Yellow River delta, with a maximum value above $200 \text{ Bq/m}^2/\text{y}$. These observations support the fact that riverine input makes important ^{137}Cs

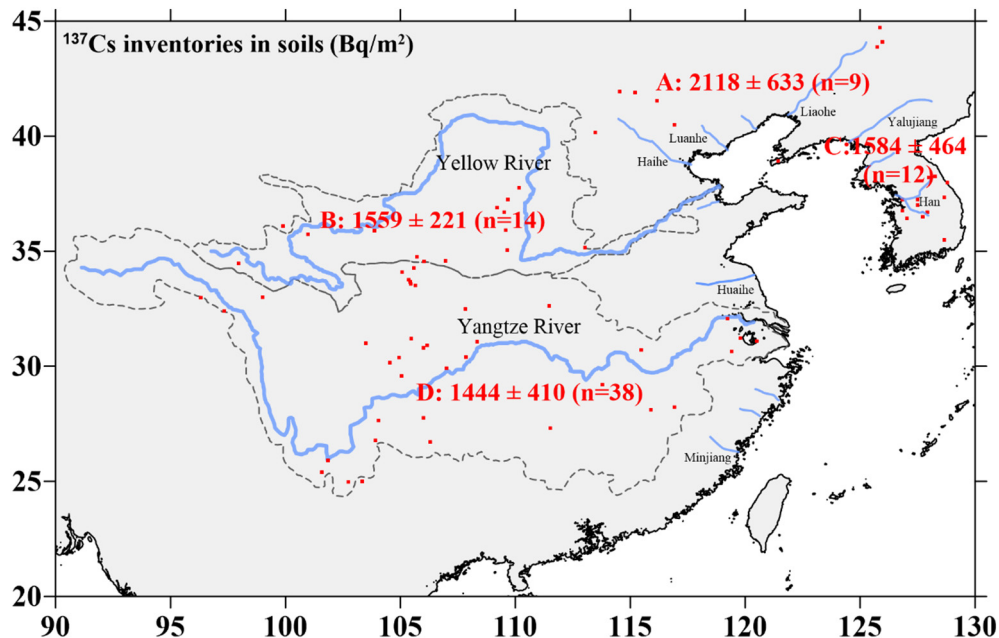


Fig. 5. The synthesis of decay-corrected ^{137}Cs inventory (Bq/m^2) in river drainage basin surrounding the ECSs. The red symbols show locations of soil cores in published literatures and the details of ^{137}Cs inventory data are summarized in Table S4 (Menzel et al., 1987; Zhang et al., 1990, 1994, 1997, 1998, 2003, 2006; Quine et al., 1992; Li et al., 1993, 1995; Lee et al., 1996; Lu and Higgitt, 2000; Cao et al., 2001; Zhang et al., 2002, 2011; Jin et al., 2004; Pu et al., 2004; Yan and Shi, 2004; Zhang et al., 2004; Hu et al., 2005; Yan and Yang, 2005; Zhao et al., 2005; Fang et al., 2006; Hua et al., 2006; Tang et al., 2006; Zheng et al., 2007; Yan et al., 2008; Li et al., 2009b; Li et al., 2009a; Zhang et al., 2009b; Jiang, 2010; Zhang et al., 2010; Shao et al., 2011; Wang et al., 2011; Yang et al., 2011; Jia et al., 2012; Meusburger et al., 2013; Liang et al., 2014; Wang et al., 2014; Zhang et al., 2014; Li et al., 2015; Xu et al., 2015; Zhang et al., 2015; Ma et al., 2016; Li et al., 2017; Su et al., 2017). For calculating the riverine input, four parts are divided and corresponding arithmetic mean value is shown. Part A: Haihe–Liaohe basin; Part B: Yellow River basin; Part C: Korean Peninsula; Part D: Yangtze River basin.

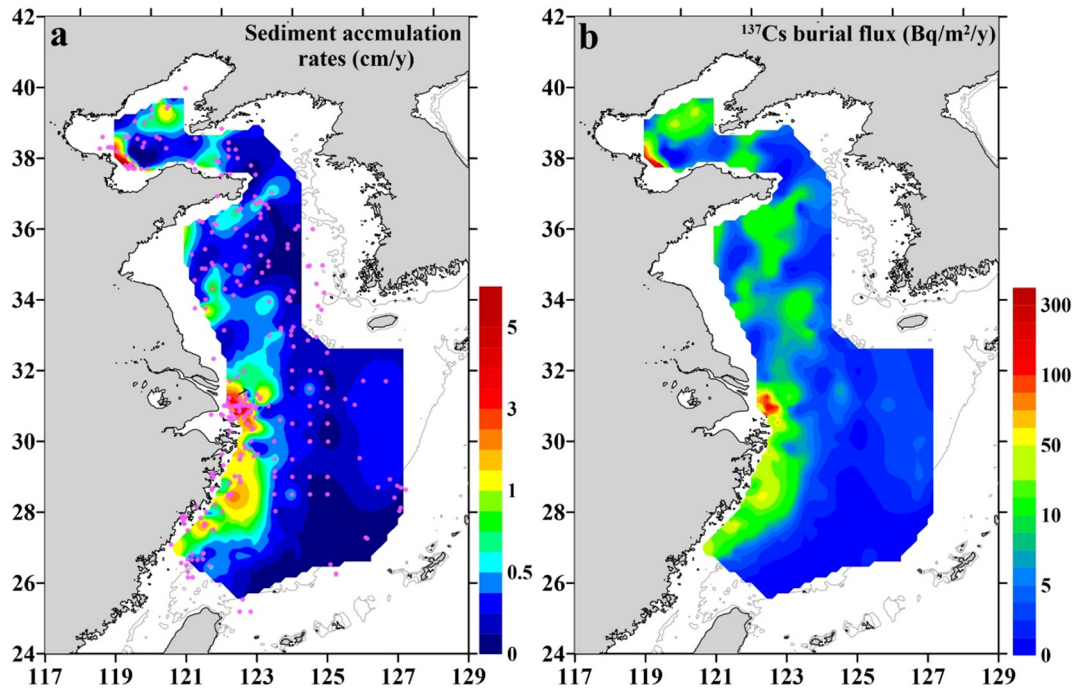


Fig. 6. Sediment accumulation rates (cm/y) and ^{137}Cs burial flux ($\text{Bq}/\text{m}^2/\text{y}$) in the ECSs. The purple symbols show locations of sediment cores in published literatures and the details of sediment accumulation rates are summarized in Table S5 (DeMaster et al., 1985; Zhao et al., 1991; Alexander et al., 1991; Yang et al., 1993; Xie et al., 1994; Chung and Chang, 1995; Huh and Chen, 1999; Huh and Su, 1999; Xia et al., 1999, 2004; Lin et al., 2000; Li et al., 2002, 2006; Su and Huh, 2002; Oguri et al., 2003; Chen et al., 2004; Zhang et al., 2005; Liu et al., 2006; Fang et al., 2007; Guo et al., 2007; Lim et al., 2007; Yang and Youn, 2007; Yang and Chen, 2007; Zhang et al., 2009a; Dong et al., 2009; Feng et al., 2009; Liu et al., 2009; Zhang et al., 2009c; Hu et al., 2011a; Gao et al., 2011; Liu and Fan, 2011; Youn and Kim, 2011; Wen, 2012; Duan et al., 2013; Li et al., 2013; Zhu et al., 2014; Liu et al., 2015; Hu et al., 2016; Zhou et al., 2016; Wang et al., 2017a; Qiao et al., 2017; Lin et al., 2018; Xu et al., 2018).

contributions in the estuary areas. The ^{137}Cs activities and sediment accumulation rates in the outer shelf of the ECS are both low, and thus, the ^{137}Cs burial flux is very low. The weighted averages of ^{137}Cs burial fluxes to surface sediments in the BS, YS and ECS are calculated to be approximately $22.4 \pm 8.7 \text{ Bq}/\text{m}^2/\text{y}$, $6.9 \pm 1.6 \text{ Bq}/\text{m}^2/\text{y}$, and $7.3 \pm 2.3 \text{ Bq}/\text{m}^2/\text{y}$, respectively. Accordingly, the total ^{137}Cs burial fluxes can be estimated to be $(7.4 \pm 2.9) \times 10^{11} \text{ Bq}/\text{y}$, $(1.3 \pm 0.3) \times 10^{12} \text{ Bq}/\text{y}$ and $(2.8 \pm 0.9) \times 10^{12} \text{ Bq}/\text{y}$ in the BS, YS and ECS, respectively. The ^{137}Cs burial fluxes in the YS and ECS are far greater than the riverine input and atmospheric fallout, which suggests that the dominant source of ^{137}Cs in the YS and the ECS is oceanic input from the WNP by the KC, although riverine input is important for the estuary areas.

4.2. ^{137}Cs budget in the ECSs

In a given region, ^{137}Cs activity in the water column is determined by external input (atmospheric fallout, riverine input and oceanic input) and output (sedimentary deposition and natural decay). Therefore, the temporal change of ^{137}Cs inventory can be expressed by a box model:

$$dI/dt = R_{\text{CS}} + A_{\text{CS}} + O_{\text{CS}} - \lambda I - B_{\text{CS}} \quad (2)$$

where I is the inventory of ^{137}Cs in the water column (Bq), λ (0.023 y^{-1}) is the natural decay constant of ^{137}Cs , and R_{CS} , A_{CS} , O_{CS} and B_{CS} refer to the riverine input (Bq/y), atmospheric fallout flux (Bq/y), net oceanic contribution and burial flux into surface sediments (Bq/y), respectively. The R_{CS} , A_{CS} and B_{CS} have been discussed above, but the temporal change of ^{137}Cs inventory (dI/dt) still needs to be solved. Under steady-state conditions, the ^{137}Cs inventory remains constant over time such that dI/dt is equal to zero. However, exponentially decreasing trends over time of ^{137}Cs activity in surface seawaters have been observed in many seas such as the Japan Sea (Kasamatsu and Inatomi, 1998) and the Pacific Ocean (Hirose and Aoyama, 2003). The historical data of

^{137}Cs in the surface water of the YS and ECS are plotted against time in Fig. 7. The ^{137}Cs data are fitted to exponentially decreasing curves, and the apparent half-lives of ^{137}Cs are estimated to be 15.1 y for the ECS and 7.7 y for the YS. The value obtained in the ECS is slightly lower than those found in the Japan Sea (18.7 y) (Kasamatsu and Inatomi, 1998), the South China Sea (16.9 y) (Wu, 2018) and the Northwest Pacific Ocean (17 y) (Hirose and Aoyama, 2003). Unfortunately, the data in the BS are so rare that we only can use the value estimated in the YS as a compromise. The ^{137}Cs inventory at each vertical profile can be calculated by integrating ^{137}Cs activities at each depth (Ito et al., 2003):

$$I = \frac{1}{2} \left\{ \sum_{i=1}^N (C_i + C_{i+1})(d_{i+1} - d_i) + 2C_1d_1 + 2C_N(d_B - d_N) \right\} \quad (3)$$

where I is ^{137}Cs inventory (Bq/m^2), N is the number of sampling depths, C_i is the ^{137}Cs activities at depth i (Bq/m^3), d_i is the i th sampling depth (m), d_1 is the first sampling depth (m), and d_B is the total depth to the bottom (m). Using this formula and surface area of the YS ($1.95 \times 10^5 \text{ km}^2$), we can calculate that the ^{137}Cs inventories in the YS is $(3.8 \pm 3.0) \times 10^{12} \text{ Bq}$. For the BS, considering its water depth is shallow, we assume the concentration of ^{137}Cs in the water profiles of the Bohai Sea is homogeneous, and thus we use the ^{137}Cs concentration in surface water to calculate that the ^{137}Cs inventories in the BS is $(3.8 \pm 2.0) \times 10^{10} \text{ Bq}$. In the ECS, we divide it into two parts (shallow waters (<100 m) and deep waters) to calculate ^{137}Cs inventories. For the shallow waters, we assume the concentration is homogeneous in vertical distribution. For the deep waters part, we use the vertical profiles to estimate the ^{137}Cs inventory. Finally, the ^{137}Cs inventory in the ECS water is calculated to be $(5.1 \pm 3.3) \times 10^{13} \text{ Bq}$ (details are shown in supplemental file). Combining the apparent half-lives and ^{137}Cs inventories in the water column, the dI/dt in the BS, YS and ECS are estimated to be $(0.3 \pm 0.2) \times 10^{10} \text{ Bq}/\text{y}$, $(3.3 \pm 2.6) \times 10^{11} \text{ Bq}/\text{y}$ and $(2.4 \pm 1.5) \times 10^{12} \text{ Bq}/\text{y}$, respectively.

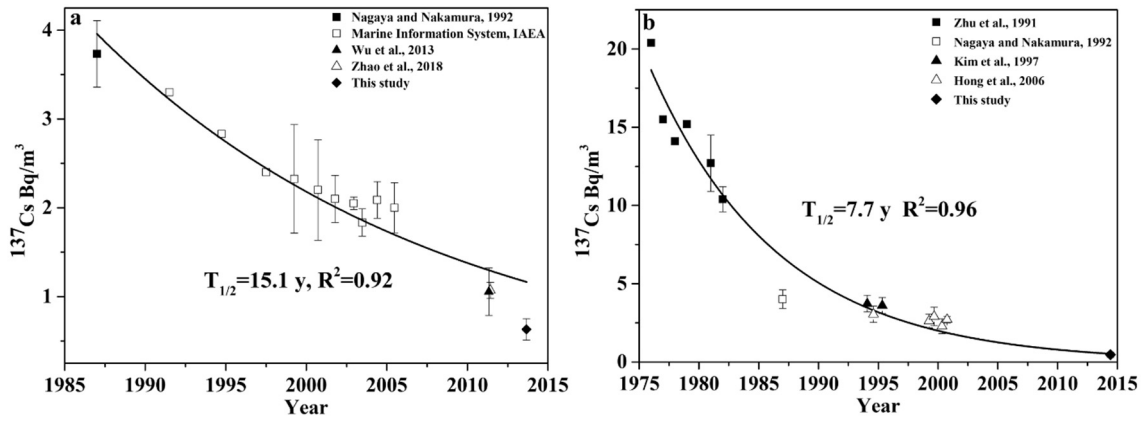


Fig. 7. Temporal changes of ^{137}Cs activities in surface seawater of the ECS (a) and YS (b) during the past few decades (Zhu et al., 1991; Nagaya and Nakamura, 1992; Kim et al., 1997; Hong et al., 2006; Wu et al., 2013; Zhao et al., 2018; MARIS data are from <http://maris.iaea.org/Home.aspx>).

Finally, using the box model, the interlinked budget balance of ^{137}Cs between the BS, YS and ECS is constructed (Fig. 8). The BS is a typical semienclosed sea and connect to the YS only through the Bohai Strait; thus, it is an ideal model for a box calculation and can be served as a starting point for budget calculation. The net oceanic input for the BS is calculated to be $(2.2 \pm 2.9) \times 10^{11}$ Bq/y, and this value actually is the ^{137}Cs outflow flux from the YS to the BS. Subtracting the outflow flux from the budget calculation of the YS, the ^{137}Cs input from the ECS to the YS is calculated to be $(1.7 \pm 0.4) \times 10^{12}$ Bq/y. Likewise, about $(7.4 \pm 1.9) \times 10^{12}$ Bq/y of ^{137}Cs should be transported from the WNP to the ECS, this value is very close to the previously estimated result (7.7×10^{12} Bq/y) obtaining by compartment model POSIDON-R (Maderich et al., 2014). Apparently, the oceanic input from the open ocean is a dominant source for ^{137}Cs in the ECSs, and the oceanic input is 5.2 times higher than the sum of the riverine input and atmospheric fallout. Considering the fact that we might overestimate the riverine input of ^{137}Cs , the result reported here should be a conservative estimation.

The KC brings approximately $7.9 \times 10^{14} \text{ m}^3 \text{ y}^{-1}$ of high salinity water to the ECS (Kagimoto and Yamagata, 1997). In this study, the mean ^{137}Cs activity in the 100–200 m layer of the KC is calculated to be $1.3 \pm 0.4 \text{ Bq/m}^3$. Multiplying the flow of the KC into the ECS and mean ^{137}Cs activity, we can roughly estimate that $(10.3 \pm 3.2) \times 10^{14}$ Bq/y of ^{137}Cs passes through the ECS. Therefore, the trapped ^{137}Cs in the ECSs accounts for $0.7\% \pm 0.3\%$ (calculated by dividing the total ^{137}Cs of 10.3×10^{14} Bq/y passing through the ECS by the net oceanic input into the ECS of 7.4×10^{12} Bq/y) of the total ^{137}Cs transported by the KC. Among the oceanic inputs of ^{137}Cs , approximately 20% can be transported into the YS. However, only 3% can reach into the BS.

4.3. Implication of ^{137}Cs use for tracing water mass movement and exchange

In the estuarine regions, ^{137}Cs can be removed from the water body by adsorbed on the clay particles due to higher SPM concentration. However, due to very low K_d value of ^{137}Cs ($<500 \text{ L/kg}$) in seawater and relatively

lower SPM concentration ($<10 \text{ mg/L}$) of shelf waters, ^{137}Cs is highly soluble and can be used to trace water mass movement in shelf waters. Though there have been many studies using ^{137}Cs as a tracer for water mass movement all over the world's oceans (e.g. Bowen et al., 1980; Aarkrog et al., 1983; Povinec et al., 2011; Aoyama et al., 2008, 2011), most studies focus on the open ocean, not marginal seas. For the sake of clarity, we summarize the salinity and temperature ranges of water masses in the ECSs and plot the ^{137}Cs activities in a T-S diagram (Fig. 9). In the ECS (Fig. 9a), different color spots show blocky distribution characteristics. Low ^{137}Cs activities were found in the Yangtze Estuary and extended to the northeast to Cheju Island due to the large amount of freshwater discharge and low ^{137}Cs wash out from the catchment area. In freshwater, ^{137}Cs is highly particle-reactive, with a K_d value of approximately 10^5 L/kg (Baskaran et al., 2015; Wang et al., 2017b), and low levels of dissolved ^{137}Cs have been found in the Yangtze River water (Du et al., 2010). The mean ^{137}Cs activity of the ECS water ($0.62 \pm 0.32 \text{ Bq/m}^3$) in summer was lower than that in our previous report ($1.07 \pm 0.09 \text{ Bq/m}^3$) during May 2011 (Zhao et al., 2018). The discrepancy was mainly derived from the influence of the seasonal variation of the CDW. Moreover, the low ^{137}Cs region in spring appeared in the southeast of the Yangtze Estuary and the Zhe-Min coast (Zhao et al., 2018). The distinct difference in the low ^{137}Cs region is in line with the dispersal path of the CDW in different seasons. During the flood season, the CDW flows toward the northeast and may extend to Cheju Island, while in the dry season, the CDW flows southward in a narrow band confined to the Zhe-Min coast (Wu et al., 2011, 2014). Thus, ^{137}Cs could respond precisely to the seasonal variation of the CDW plume.

High ^{137}Cs activities were observed in the Kuroshio surface water (KSW) and the Kuroshio subsurface water (KSSW), which supports the fact that the oceanic input from the WNP is a dominant source of ^{137}Cs in the ECS. The residence time of the ECS shelf water has been proven to be $1.3 \pm 0.27 \text{ y}$ (Wang et al., 2018c). Thus, the material exchange between the ECS and the WNP through shelf-edge intrusion of the KC is active (Wong et al., 2000). The massive intrusion mainly occurred northeast of Taiwan Island due to topographic deflection and is

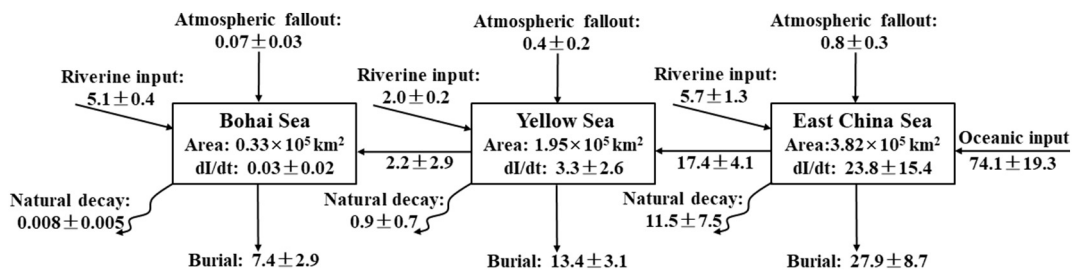


Fig. 8. Interlinked budget balance of ^{137}Cs (10^{11} Bq/y) between the BS, YS and ECS.

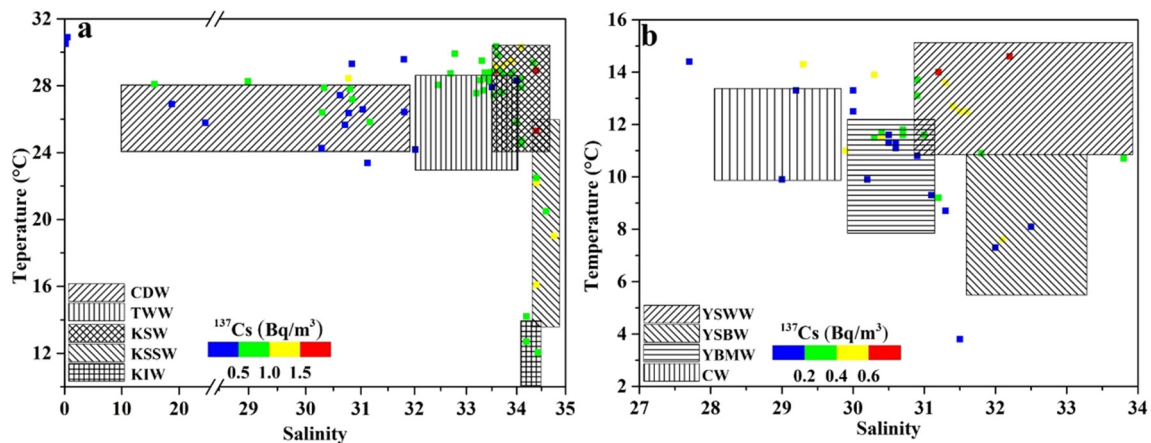


Fig. 9. ^{137}Cs activities of seawater samples plotted in T-S diagram. Different colors represent ^{137}Cs activities, note difference in scale of color bar. The temperature and salinity range of water masses in the ECSs is from previous study (Liu et al., 1993; Hur et al., 1999; Chen, 2009; Qi et al., 2014). CDW: Changjiang Diluted Water; KSW: Kuroshio Surface Water; KSSW: Kuroshio Subsurface Water; KIW: Kuroshio Intermediate Water; TWW: Taiwan Warm Water; YSWW: Yellow Sea Warm Water; YSBW: Yellow Sea Bottom Water; YBMW: Yellow-Bohai Sea Mixed Water; CW: Coastal water.

a year-round phenomenon, characterized by the upwelling of the KSSW (Liu et al., 1992; Wong et al., 2000). In addition, small-scale intrusions in the form of frontal eddies could occur along the ECS shelf edge (Yanagi et al., 1998). Therefore, the intrusion of the KSSW is a main mechanism of ^{137}Cs import into the ECSs. The ^{137}Cs vertical profiles near the shelf edge further confirms our hypothesis (Fig. 3c). Among five stations located along the ECS shelf edge, the maximum values were observed in the subsurface layer at a water depth of 100–200 m in four vertical profiles of ^{137}Cs (F8, D10, C12 and C13). On the basis of numerical simulation and field observations, Yang et al. (2012) suggested that the KSSW could intrude near the Yangtze river mouth via the bottom layer. Similar results were also obtained using $\delta^{18}\text{O}$, δD and $^{129}\text{I}/^{127}\text{I}$ as tracers (Lian et al., 2016; Wang et al., 2018a). In this study, at station C2, the ^{137}Cs level in the bottom layer was approximately five times higher than that in the surface layer. The high ^{137}Cs concentration in the bottom layer near the Yangtze River mouth is a clear signal of the intrusion of the KSSW. In the north of the ECS, the relatively high ^{137}Cs level showed a mixing between the KC/TWC waters and the CDW water. This result indicates that a portion of ^{137}Cs could be transported to the YS from the ECS via the YSWC, which is consistent with the budget calculation as discussed above.

Regarding the YS-BS (Fig. 9b), relatively high ^{137}Cs activities were observed in the Yellow Sea Warm Water (YSWW). The high ^{137}Cs level decreased sharply northward, but smoothly westward (Fig. 2b). Such a distribution pattern suggests that the YSWC does not exactly flow along the YS trough but is displaced significantly westward. The westward shift of the YSWC has already been pointed out in several previous studies (Xie et al., 2002; Huang et al., 2005; Ma et al., 2006; Lin et al., 2011). The warm water intrusion can reach approximately 122.5°E at 35°N ; then, a large portion turns to northeast, but a small warm water tongue continues trending northwestward to the Shandong coast (Lin et al., 2011). The BSCC carries extremely low ^{137}Cs water into the YS and then meets the YSWW. Thus, the lower ^{137}Cs south of the Shandong Peninsula might be a result of mixing between the YSWC water and the BSCC water. According to our mass balance calculation, only 4% of ^{137}Cs is derived from oceanic input, therefore, the lower ^{137}Cs activity in the BS surface waters should be attributed to the low concentration in the Yellow River and other rivers. In addition, the spatial distribution of ^{137}Cs in the surface water with low values in May 2014 suggests that in spring, the YSWC only intrudes into the southern YS and eastern part of the northern YS but not into the BS. The 4% of oceanic input of ^{137}Cs to the BS should be contributed by YSWC transport in winter (Guan, 1994).

Overall, the spatial and vertical distribution of ^{137}Cs is significantly controlled by the complex current system in the ECSs. The high level of ^{137}Cs in oceanic water masses and low level of ^{137}Cs in riverine and coastal waters make ^{137}Cs a good indicator for tracing water mass movement and interaction in the ECSs. It has been reported that KSSW is rich in nutrients, especially phosphate (Liu et al., 2000; Zhang et al., 2007); thus, the results in this study are expected to help us to understand nutrient transport and budgets in the ECSs.

4.4. Implication of ^{137}Cs as a tracer for sediment transport in river-dominated margins

In freshwater, ^{137}Cs is preferentially adsorbed by clay minerals (especially illite) and is strongly associated with cation-exchange sites (Duran et al., 2004). Once these clay particles were discharge to the sea, the desorption of ^{137}Cs from particles was reported to be small, with a maximum value of 9% (Takata et al., 2015). Ohnuki and Kozai (2013) investigated the desorption of ^{137}Cs and found that it was <20% even when leaching with 1 M HCl. Thus, ^{137}Cs in clay matrices is nearly immobile and can be transported in pace with the terrestrial fine sediments in river-dominated marginal seas. The sediments from the Yangtze River and Yellow River are abundant with fine particles, and mean grain size (φ) of suspended particles in the Datong station and Lijin station were documented to be 6.64 and 5.72, respectively (CWRC, 2014). In addition, X-ray diffraction analysis demonstrated that illite is the most abundant clay mineral in both sediments from the Yangtze River and the Yellow River (~70% and ~60%, respectively) (Fan et al., 2001; Wang and Yang, 2013). Thus, the majority of suspended sediments, especially fine particles, originating from the Yangtze River and Yellow River are tagged with ^{137}Cs . These ^{137}Cs -tagged sediments deposit predominantly in river mouths at first, and subsequently, a large portion of deposited sediments is eroded, and re-suspension then disperses throughout the entire ECSs by following the coastal currents.

In the Yangtze River mouth, the difference between short- and long-term sedimentation rates, 4 cm/month versus 1–5 cm/year, suggests that a large proportion of sediment are eroded on a seasonal basis and transported farther offshore (McKee et al., 1983) that corresponds to the southward alongshore dispersion of fine sediments as revealed by the low mean grain size and high ^{137}Cs activity. The outer shelf of the ECS was covered by coarse sand due to the relict sediment deposition during the last glacial period (Qin et al., 1996), and thus, the ^{137}Cs in the outer shelf is lower. Near the shelf edge of the ECS, relatively high

activities are also observed. Previous studies have revealed that suspended sediments can be transported from the inner shelf to the shelf edge in the autumn and winter (Yanagi et al., 1996; Iseki et al., 2003; Wang et al., 2018b). The near-bottom transport is the dominant mechanism of sediments across the shelf, as proven by the total suspended material profile along the PN section (Deng et al., 2006). Therefore, relatively high activities near the shelf edge of the ECS are a result of the across-shelf transport of sediments. The extremely high ^{137}Cs levels with a decreasing trend southward coincide with the southward dispersion of fine particles discharged from the Yellow River. Fine sediments from the Yellow River can travel via the BSCC through the Bohai Strait into the YS. Due to the cyclonic gyre upwelling/downwelling, most of these sediments are deposited in the central Yellow Sea (Lee and Chough, 1989; Alexander et al., 1991; Lim et al., 2007; Yang and Youn, 2007). Previous studies suggested that the fine sediments in the southwestern Cheju Island mud area are derived from the Yangtze River, Yellow River, and erosion of the old Yellow River delta along the north Jiangsu coast (Milliman et al., 1985; Alexander et al., 1991; Liu et al., 2003; Lim et al., 2007). The sediments from the Yangtze River and the Yellow River both have high ^{137}Cs activities, but ^{137}Cs in the old Yellow River delta is presumably low because there was no ^{137}Cs before 1855. If contributions from the three sources are approximately the same, the ^{137}Cs activity should be relatively high. Therefore, the low ^{137}Cs in the southwestern Cheju Island mud area might indicate that it was mainly originated from the old Yellow River delta, but not from the Yangtze River and Yellow River, at least after 1960's.

The ^{137}Cs activity in surface sediments is plotted against mean grain size (Fig. 10), that shows a good correlation. In the Kara Sea of the Russian Arctic, a positive correlation was also observed (Baskaran et al., 1996). The good correlation between ^{137}Cs activity and mean grain size further indicates that ^{137}Cs can be used as an effective tracer to track the dispersal pathways of fine sediments derived from the Yangtze River and Yellow River, even in a very complex and dynamic system. It has been reported that riverine input sediments can carry many other particle-active pollutants. Lin et al. (2002) pointed out that heavy metals, such as Fe, Zn and Cu, are easily adsorbed by fine sediments. Hu et al. (2011b) found positive correlations between mean grain size and concentrations of dichlorodiphenyltrichloroethanes (DDTs) and hexachlorocyclohexanes (HCHs) in surface sediments in the YS. Therefore, the particle effect of ^{137}Cs is not only helpful to track the dispersal pathways of fine sediments but also useful to understand the transport processes and ultimate fates of other particle-reactive pollutants.

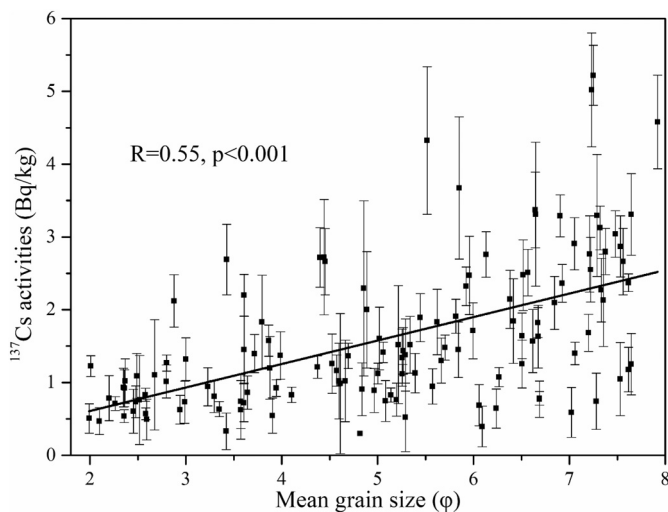


Fig. 10. The correlation diagram between mean grain size (ϕ) and ^{137}Cs activities (Bq/kg) of surface sediments.

5. Conclusion

We have measured ^{137}Cs activity in the water column and surface sediments of the ECSs. At the same time, the ^{137}Cs inventory data from land surfaces as well as sediment accumulation rates have been collected to evaluate riverine input and ^{137}Cs burial. ^{137}Cs activities in the ECSs water ranged from 0.03 to 1.92 Bq/m³ with high ^{137}Cs observed in the shelf edge areas and low ^{137}Cs observed in the estuary and coastal areas. In surface sediments, ^{137}Cs activities ranged from 0.30 to 5.22 Bq/kg, with a high activity observed in the Zhe-Min coast and the central YS mud area. There was no ^{134}Cs signal in any samples, suggesting that the impact of the Fukushima accident on the ECSs was very limited. Using the ^{137}Cs inventory data of drainage basin and sediment accumulation rates over the ECSs, the total riverine inputs and ^{137}Cs burial fluxes into surface sediments are estimated to be $(1.3 \pm 0.1) \times 10^{12}$ Bq/y and $(4.8 \pm 1.0) \times 10^{12}$ Bq/y. Mass balance calculations show that the oceanic input from the open ocean is a dominant source for ^{137}Cs in the ECSs that is 5.2 times higher than the sum of the riverine input and atmospheric fallout.

The vertical profiles of ^{137}Cs with a maximum at a water depth of 100–200 m in the continental shelf edge and much higher ^{137}Cs in bottom waters compared to surface waters in the Yangtze Estuary reveal that the upwelling of the Kuroshio subsurface water is the main mechanism of ^{137}Cs import into the ECSs. The spatial distribution of ^{137}Cs suggests that in spring, the YSWC can intrude into the southern YS and east part of the northern YS but not into the BS. The high level of ^{137}Cs in oceanic water masses and low level of ^{137}Cs in riverine and coastal waters make ^{137}Cs a good indicator for tracing water mass movement and interaction in the ECSs. The fine grain size and low ^{137}Cs of surface sediments in the southwest region of Cheju Island suggest that the source of these regional sediments is dominated by old Yellow River delta input. There is a good correlation between ^{137}Cs activity and mean grain size (ϕ) in surface sediments, suggesting that ^{137}Cs can serve as an effective tracer to track dispersal pathways of fine sediments and associated particle-reactive pollutants in river-dominated marginal seas.

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Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.scitotenv.2019.04.001>.

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